

1       **Recommendations to derive quality standards for chemical pollutants in reclaimed**  
2                               **water intended for reuse in agricultural irrigation**

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20 **Abstract** The reuse of treated municipal wastewater (herein referred to as reclaimed water) in  
21 agricultural irrigation (RWAI) as a means to alleviate water scarcity is gaining increasing policy  
22 attention, particularly in areas where water demand mitigation measures have proved  
23 insufficient. However, reclaimed water reuse in practice is lagging behind policy ambition, with  
24 <2.5% of it reused in a European context. A key barrier identified as limiting its full valorisation  
25 is concern over its impact on human and environmental health. To address this concern, and to  
26 meet further objectives including achieving parity between current reclaimed water reuse  
27 guidelines operational in various Member States, the European Commission has proposed a  
28 regulation which identifies minimum quality requirements (MQR) for a range of  
29 microbiological and physico-chemical parameters but the inclusion of compounds of emerging  
30 concern (CECs) in terms of the determination of quality standards (QS) is missing. This paper  
31 reviews the existing pertinent EU legislation in terms of identifying the need for CEC QS for  
32 RWAI, considering the scope and remit of on-going pan-European chemicals prioritisation  
33 schemes. It also evaluates opportunities to link in with the existing EQS derivation methodology  
34 under the EU WFD to address all protection targets in the environmental compartments exposed  
35 via potential pathways of RWAI. Finally, it identifies the main data gaps and research needs for  
36 terrestrial ecosystems, the removal efficiency of CECs by WWTPs and transformation products  
37 generated during the wastewater reuse cycle.

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## 40 **1. Introduction**

41 Over 2 billion people live in countries experiencing high water stress, with approximately 4  
42 billion people experiencing severe water scarcity during at least one month of the year (WWAP,  
43 2019). Whilst freshwater is relatively abundant in the European Union (EU), around 30% of  
44 the total European population, experienced water scarcity conditions in the summer of 2015

45 compared to 20% in 2014 (EEA, 2019). Whilst total water abstractions within Europe have  
46 decreased by an estimated 19% since 1990 (driven by efficiency gains likely to continue in the  
47 coming years), water stress hotspots are predicted to remain and even grow given continued  
48 pressures such as expanding urbanisation, increasing population and climate change (EEA,  
49 2019). Forecasts such as these highlight the urgent need to utilise alternative water resources  
50 such as treated wastewater (referred to henceforth as reclaimed water reuse) when water  
51 efficiency, demand management and improved agricultural practices are not sufficient to  
52 prevent water scarcity.

53 A study by WWAP (2017) distinguishes three main types of reclaimed water reuse: 1) direct  
54 potable reuse; 2) indirect potable reuse; and 3) reuse for non-drinking purposes including  
55 agricultural irrigation. This study reports that the reclaimed water reuse in agriculture is an area  
56 of great potential and indeed, the European Environment Agency (EEA) report on European  
57 waters indicates that in the Spring of 2014, the agricultural sector used 66 % of the total water  
58 used in Europe (EEA, 2018). Whilst current EU legislation encourages reclaimed water reuse  
59 through the Urban Waste Water Treatment Directive (UWWTD, EEC 1991) and the Water  
60 Framework Directive (WFD, EU 2000), these legislative pieces only refer to reuse practices in  
61 brief without specifying conditions for reclaimed water quality for reuse practices. During the  
62 2012 fitness check on EU freshwater policy (EC, 2012), industry stakeholders raised concerns  
63 about the lack of EU quality standards for reclaimed water reused in agricultural irrigation  
64 (RWAI), with potential impacts on the free movement of agricultural produce in the single  
65 market highlighted. Six Member States (MSs) (Cyprus, France, Italy, Greece, Spain and  
66 Portugal) have requirements in place which set-out quality requirements for reclaimed water  
67 reuse either in legislation or as non-regulatory standards, but these requirements vary  
68 significantly in terms of both parameters included and their associated values. Additionally, it  
69 should be noted that EU citizens in many MSs have expressed their concerns about water safety

70 e.g. the European Citizens' Initiative "Right2Water", ECI R2W, 2012). and many international  
71 water reuse initiatives like in the US or Singapore have faced public opposition (Voulvoulis,  
72 2018).

73 In response to identified concerns over variations in quality standards and associated  
74 implications for the transnational shipment of irrigated crops, the European Circular Economy  
75 Action Plan (EC, 2015) announced activities to facilitate reclaimed water reuse, including a  
76 proposal to develop legislation on minimum quality requirements (MQRs) for RWAI and  
77 groundwater recharge. In May 2018, the European Commission (EC) put forward a legislative  
78 proposal for a regulation of the European Parliament and of the Council on MQRs for reclaimed  
79 water reuse seeking to incentivise reuse, while ensuring a high level of protection of health and  
80 the environment (EU, 2018). Reclaimed water, defined as urban wastewater that has undergone  
81 treatment in a reclamation facility, will be used to irrigate food crops, processed food crops and  
82 non-food crops. The EC estimates that the proposal could enable the reuse of more than half of  
83 the current volume of water coming from EU wastewater treatment plants within irrigation,  
84 resulting in a reduction of water stress of >5%. Despite requesting over 400 amendments to the  
85 proposed text, the European Parliament adopted its first reading position on 12 February 2019  
86 and in the Council, the proposal is being examined by the Working Party on the Environment.  
87 The main political issues that have emerged from the discussion include the degree of flexibility  
88 the EU instrument should offer to MSs and the stringency of the MQRs for reclaimed water  
89 quality (EPRS, 2018). The regulation proposal defines MQRs for microbiological (e.g.  
90 *Legionella*, *E. coli*) and physicochemical (BOD, TSS and turbidity) parameters. It states that  
91 competent authorities would have the possibility to impose additional requirements, based on a  
92 risk management plan submitted by the reclamation plant operator, or on the need to mitigate  
93 unacceptable risks to health or the environment. The legal proposal is based on a review of  
94 current knowledge pertaining to reclaimed water reuse developed by EC Joint Research Centre

95 (JRC, 2017). The development of the proposal adopted a tiered approach, requesting reviews  
96 of proposed MQRs from the EU Scientific Committee on Health, Environment and Emerging  
97 Risks (SCHEER) and the European Food Safety Authority (EFSA). Both (EFSA, 2017 and  
98 SCHEER, 2017) were of the opinion that the proposal provided insufficient protection to  
99 environmental and human health especially with regard to Compounds of Emerging Concern  
100 (CECs) and disinfection by-products associated with wastewater treatment plants (WWTPs).  
101 SCHEER recommended developing common criteria and detailed guidance on MQRs for  
102 priority CECs and EFSA identified the need to assess their impact on human, animal and  
103 environmental health. In addition, the JRC presented their findings and recommendations at  
104 several public events and scientific meetings. On various occasions, members of the COST  
105 Action ES1403 NEREUS also presented the current stage of knowledge concerning CECs and  
106 wastewater reuse in such meetings. The COST Action ES1403 established a multi-disciplinary  
107 network to determine which of the current challenges related to wastewater reuse are the most  
108 concerning in relation to public health and environmental protection, and how these could be  
109 overcome. A core activity of the network was the collaborative development of a framework to  
110 support the qualitative assessment of risks associated with reclaimed water reuse. At the final  
111 conference of the NEREUS COST Action in Limassol, Cyprus, in October 2018, a panel was  
112 organised to discuss the ‘big unknowns concerning safe and sustainable wastewater reuse’ and,  
113 more specifically, how effective the proposed MQRs are. The panel concluded that, in their  
114 current form, the proposed MQRs provided insufficient protection both to environmental and  
115 human health. This international discussion panel has been the starting point for a more in-  
116 depth consideration on how to address this identified knowledge gap on chemical pollutants,  
117 leading to the development of practical recommendations on how to prioritise substances and  
118 develop MQRs to support the safe use of RWAI.

119

## 120 **2. Identification and prioritisation of chemical contaminants of concern**

121 Several studies in the literature report the presence of a range of organic and inorganic  
122 contaminants in the effluents of urban WWTPs in the EU. For example, Aguayo et al (2004)  
123 identified more than 49 compounds in the organic fraction of effluents from seven urban  
124 WWTPs in close proximity to urban and industrial areas within the community of Madrid  
125 (Spain). Karvelas et al. (2003) investigated the occurrence and the fate of heavy metals (Cd,  
126 Pb, Mn, Cu, Zn, Fe and Ni) in the urban WWTP of the city of Thessaloniki (northern Greece)  
127 and found that 47–63% of influent concentrations of Cd, Cr, Pb, Fe, Ni and Zn remain in the  
128 treated effluent. According to Rizzo et al. (2019) urban WWTPs are not designed to remove  
129 CECs and secondary (e.g. conventional activated sludge process, CAS) and tertiary (such as  
130 filtration and disinfection) treatment processes are not effective in the removal of most of the  
131 CECs entering urban WWTPs. In order to identify which are the relevant substances that could  
132 be present in reclaimed water, we propose to identify the different sources of wastewater in the  
133 water reuse system as recommended in the Water Reuse Risk Management Plan defined in the  
134 Annex II of the EC reclaimed water reuse regulation proposal (EU, 2018).

135

### 136 **2.1 Identification of candidate substances reaching urban WWTPs**

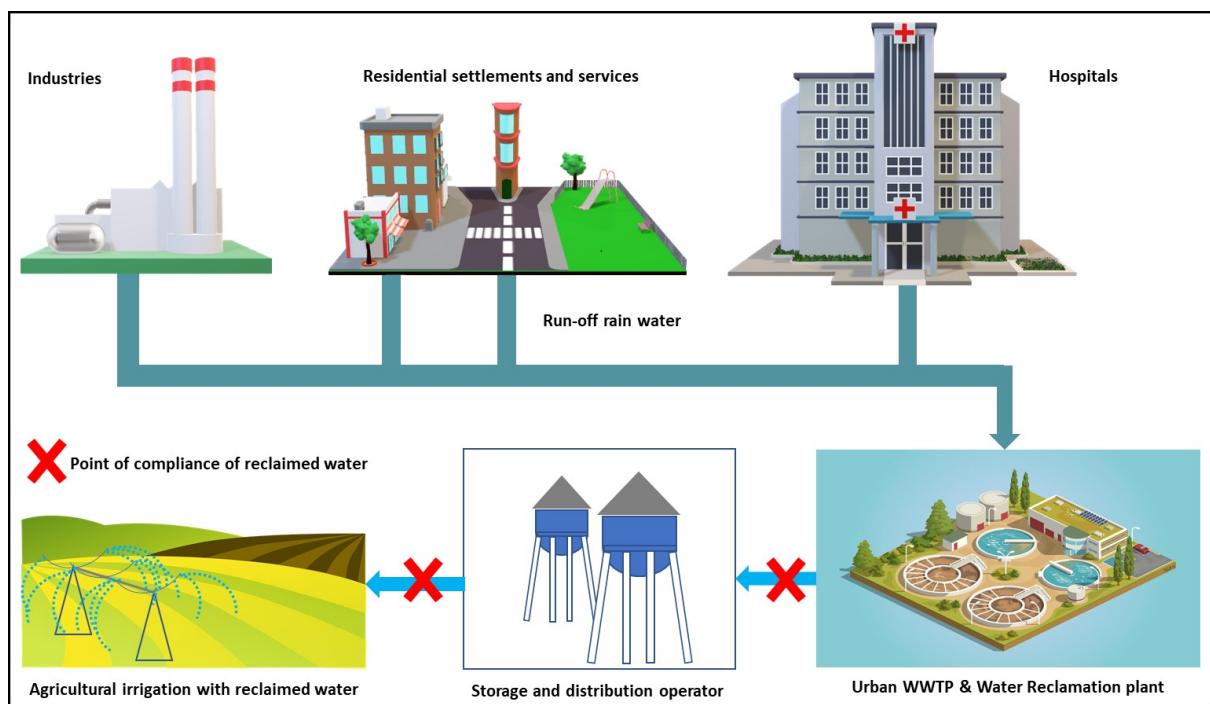
#### 137 *2.1.1 Sources of wastewater and European policy landscape*

138 The EC proposal (EU, 2018) focuses on treated wastewater as covered by the UWWTD (EEC,  
139 1991) where urban wastewater is defined as domestic wastewater or the mixture of domestic  
140 wastewater with industrial wastewater and/or run-off rainwater (see Fig. 1). Domestic  
141 wastewater refers to effluents from residential settlements and services which originate  
142 predominantly from the human metabolism and from household activities. All substances  
143 contained in consumer products but also professionals' products related to consumer services  
144 could be released into such wastewater. Industrial wastewater refers to discharges from

145 premises used to carry out a trade or industry as described in Annex I of the Industrial Emissions  
146 Directive (IED; EU, 2010). All substances manufactured, formulated or used by the referenced  
147 categories of activities could be released in such wastewater. The UWWTD states that the  
148 discharge of industrial wastewater to urban WWTPs is subject to prior authorizations by MSs  
149 and sets the requirement to both protect the health of staff and ensure that discharges from these  
150 plants do not adversely affect the environment. However, the effectiveness of this Directive  
151 almost 25 years after its adoption is now questioned and its evaluation was recently initiated by  
152 the EC, to consider in particular to which extent its quality standards (both in relation to  
153 pollutants listed and limit values identified), reflect technological developments and meet  
154 today's challenges. Industry's wastewater releases into water bodies (direct releases) as well as  
155 wastewater releases into public sewage which ends within urban WWTP (indirect releases) are  
156 among the key aspects regulated by the IED (EU, 2010). All permit conditions must be based  
157 on the environmental protection level provided by the use of Best Available Techniques (BAT)  
158 with associated emission limit values identified for each installation. However, according to the  
159 BAT for common wastewater and waste gas treatment/management systems in the chemical  
160 sector (BAT, 2016), the main risk to be addressed when discharging industrial effluents to an  
161 urban WWTP is to ensure that pollution levels in the effluent will not damage or diminish sewer  
162 system performance. In order to reduce emissions to water, BAT involves the pre-treatment of  
163 wastewater that contains pollutants that cannot be dealt with adequately during final wastewater  
164 treatment. Associated emission levels to water (BAT-AELs) are recommended for indirect  
165 releases however, they are set for groups of chemicals (i.e. some heavy metals and adsorbable  
166 organically bound halogens) and not for individual substances making difficult to assess their  
167 efficacy to control the risk of prioritised substances identified for RWAI. Finally, the European  
168 Pollutant Release and Transfer Register Regulation (E-PRTR; EU, 2006) places a legal  
169 obligation on the EC and the MSs to establish a coherent, EU-wide pollutant register concerning

170 emissions from industrial activities, including WWTPs. The E-PRTR is the largest industrial  
 171 emissions database in Europe, containing data on more than 90 substances from 45 economic  
 172 sectors. However, it is recognised that the current scope of the E-PRTR does not capture all  
 173 pertinent industrial emissions to water as the substances covered have not been revised since  
 174 the regulation was adopted and reporting is required only for certain activities, emission  
 175 thresholds and urban WWTPs serving greater than 100 000 population equivalents (EEA,  
 176 2019). Stormwater run-off is not referred to in the UWWTD but, depending on land use and  
 177 weather conditions, can be an important additional source of pollutants entering urban WWTPs  
 178 (e.g. traffic-related activities, combustion products) as described by Lundy et al. (2011),  
 179 Christoffels and al. (2016) and Brudler et al. (2019).

180



181

182 **Figure 1. Sources of wastewater to urban WWTP and reclaimed water cycle for reuse in**  
 183 **agricultural irrigation.**

184



185 The management of industrial and urban wastewaters is also regulated indirectly by the WFD  
186 and the Environmental Quality Standards (EQSs) Directive (EU, 2008) amended by the Priority  
187 Hazardous Substances Directive (EU, 2013) which aim to ensure that all aquatic ecosystems  
188 achieve 'good chemical and ecological status'. However, the good chemical status of water  
189 bodies is defined in these Directives as compliance with quality standards established at EU  
190 level for only 45 Priority Substances (PSs) and certain other pollutants (including ubiquitous  
191 PBTs: persistent, bioaccumulative and toxic chemicals) and hence does not comprehensively  
192 address all pertinent CECs. In addition, the WFD establishes the principles to be applied by the  
193 MSs to develop EQSs for specific pollutants that are 'discharged in significant quantities' as  
194 forming part of the assessment of ecological status and hence provides an opportunity to address  
195 pertinent CECs at national level.

196

### 197 *2.1.2 Potential chemical categories*

198 Based on the identified sources of wastewater in urban WWTPs, the chemical categories that  
199 could potentially be present in wastewater streams arriving at a treatment and reclamation plant,  
200 the corresponding chemical legislation at an EU level and the European agencies responsible  
201 for their implementation have been identified (see Table 1). It must be noted that disinfection  
202 agents used at WWTPs and releasing by-products of growing concern are biocidal products  
203 regulated by the Biocidal Products Regulation (BPR).

204

205 **Table 1: Chemical categories of compounds that can reach urban WWTPs with**  
206 **corresponding EU legislation and implementing authorities.**

<b>Chemical category</b>	<b>EU legislation</b>	<b>EU implementing authorities</b>
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Industrial chemicals	<sup>a</sup> REACH Regulation <sup>b</sup> CLP Regulation	European Chemicals Agency (ECHA)
Biocidal Products also called Biocides	<sup>c</sup> BPR Regulation <sup>b</sup> CLP Regulation	European Chemicals Agency (ECHA)
Human medicinal products also called Pharmaceuticals	<sup>d</sup> Medicinal products for human use Directive	European Medicines Agency (EMA)
Veterinary medicinal products	<sup>e</sup> Veterinary medicinal products Directive	European Medicines Agency (EMA)
Plant Protection Products (PPPs) also called Pesticides	<sup>f</sup> PPP Regulation  <sup>b</sup> CLP Regulation	European Food Safety Authority (EFSA)  European Chemicals Agency (ECHA)
Cosmetic products	<sup>g</sup> Cosmetics products Regulation	European Chemicals Agency (ECHA)
Food and Feed additives	<sup>h</sup> Food additives Regulation <sup>i</sup> Feed additives Regulation	European Food Safety Authority (EFSA)
Other substances like combustion products unintentional by-products of industrial processes (e.g. dioxins and furans)	No global legislation but some are covered by the <sup>j</sup> POPs Regulation	European Chemicals Agency (ECHA) for POPs

207 <sup>a</sup> Regulation (EC) No 1907/2006 of the European Parliament and of the Council of 18 December 2006 concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals.

209 <sup>b</sup> Regulation (EC) No 1272/2008 of the European Parliament and of the Council of 16 December 2008 on Classification, Labelling and Packaging of substances and mixtures.

211 <sup>c</sup> Regulation (EU) No 528/2012 of the European Parliament and of the Council concerning the making available on the market and use

212 of Biocidal Products.  
213 <sup>d</sup> Directive 2001/83/EC of the European Parliament and of the Council of 6 November 2001 on the Community code relating to medicinal  
214 products for human use.  
215 <sup>e</sup> Directive 2001/82/EC of the European Parliament and of the Council of 6 November 2001 on the Community code relating to veterinary  
216 medicinal products.  
217 <sup>f</sup> Regulation (EC) No 1107/2009 of the European Parliament and of the Council of 21 October 2009 concerning the placing of plant protection  
218 products on the market and repealing Council Directives 79/117/EEC and 91/414/EEC.  
219 <sup>g</sup> Regulation (EC) No 1223/2009 of the European Parliament and of the Council of 30 November 2009 on cosmetic products.  
220 <sup>h</sup> Commission Regulation (EU) No 1130/2011 of 11 November 2011 amending Annex III to Regulation (EC) No 1333/2008 of the European  
221 Parliament and of the Council on food additives by establishing a Union list of food additives approved for use in food additives, food enzymes,  
222 food flavourings and nutrients.  
223 <sup>i</sup> Regulation (EC) No 1831/2003 of the European Parliament and of the Council of 22 September 2003 on additives for use in animal nutrition  
224 <sup>j</sup> Regulation (EC) No 850/2004 of the European Parliament and of the Council of 29 April 2004 on persistent organic pollutants and amending  
225 Directive 79/117/EEC.

226  
227 The European Chemical Agency (ECHA) is the driving force among regulatory authorities in  
228 implementing the EU chemical legislation covering the majority of registered substances that  
229 are on the EU market. However, some substances considered as CECs are regulated by other  
230 European agencies e.g. pharmaceutical compounds regulated by EMA.

231  
232 **2.2 Prioritisation of candidate substances**  
233 Once the relevant chemical categories are identified, it is important to prioritise the  
234 corresponding (or candidate) substances in order to select those of higher concern and to  
235 develop a manageable list of substances for the risk management of RWAI.

236  
237 *2.2.1 Identification of potential hazards and exposure*  
238 Currently, ECHA is performing an “integrated selection and priority setting” exercise using  
239 screening methods to identify ‘Substances of Very High Concern’ (SVHCs) in the EU using  
240 information from the REACH registration (22,023 unique substances by March 2019) and the  
241 CLP (147,549 substances notified by February 2019 and 4,264 with Harmonised Classification

242 and Labelling by March 2019) databases as well as external information sources (e.g. scientific  
243 literature, online chemical databases). The objective of this screening is to identify which of the  
244 potentially hazardous substances have a high potential for exposure to humans or the  
245 environment by combining selected hazard data with use and exposure information as  
246 following:

- 247 • Persistency, Bioaccumulation and Toxicity properties (PBT/vPvB)
- 248 • Carcinogenicity, Mutagenicity, Reprotoxicity (CMR) cat 1A/1B
- 249 • Endocrine disruption (ED)
- 250 • Sensitisation
- 251 • High tonnage for wide dispersive uses

252 This exercise will contribute to implementation of the SVHC Roadmap, which aims to have all  
253 relevant currently known SVHCs included in a Candidate List by 2020 with a view to having  
254 full clarity on all registered substances by 2027. The output of this ECHA initiative will be  
255 highly relevant in the context of setting MQR for RWAI. However, it should also be noted that  
256 the following categories of chemicals, that are potentially relevant for reclaimed water reuse,  
257 are not included in the SVHC roadmap: medicinal product substances, food or feeding additives  
258 and unintentional by-products of industrial processes (except persistent organic pollutants;  
259 POPs). This is because these chemicals are not within ECHA's scope due to REACH  
260 registration and CLP notification exemptions. Moreover, PBT/vPvB and ED effects are not  
261 currently a classification criterion under the CLP regulation, which currently focuses on  
262 environmental safety from the perspective of aquatic ecosystems only. Therefore, for the  
263 purpose of identifying and prioritising substances for RWAI, it will be necessary to complement  
264 the ECHA list with further prioritisation exercises on the relevant excluded chemical categories  
265 and selection criteria.

266

267 *2.2.2 Declassification criteria: treatment steps and technologies at the urban wastewater*  
268 *treatment and reclamation plants*

269 In a second tier, prioritised substances could be declassified based on the reported treatment  
270 efficacy of the specific substance at urban WWTPs and reclamation plants. Wastewater  
271 treatment processes are generally referred to as primary (physical process eliminating mainly  
272 visible material), secondary (biological process removing organic matter through the use of  
273 microorganisms), tertiary (chemical process removing nitrogen and phosphorus), disinfection  
274 (for removing pathogens) and advanced (removing micropollutants) processes.  
275 Substance/treatment-specific removal efficiencies should be estimated and reported for each  
276 prioritised substance. In principle, removal efficiencies are available for all substances  
277 identified within drinking water standards (e.g. EU Drinking Water Directive EU, 1998 and the  
278 World Health Organization drinking water standards WHO, 2017). Gorito et al (2017) review  
279 several studies on removal performances by constructed wetlands for 24 PSs, 2 other substances  
280 with EQS as well as 8 CECs on the watch list of substances pursuant to the EQS Directive (EU,  
281 2008). For pharmaceuticals, substance/treatment specific removal efficiencies have been  
282 investigated as, for example, by primary and secondary clarifiers, bioreactors and sorption to  
283 primary sludge (Stasinakis et al., 2013) or by photocatalytic degradation (Paredes et al., 2019).  
284 Rizzo et al. (2019) critically reviewed the best available technologies for the advanced treatment  
285 of urban wastewater including consolidated (ozonation, activated carbon and membranes) and  
286 new advanced treatment methods (mainly advanced oxidation processes) analysing their  
287 efficiency in the removal of CECs. In addition, Krzeminski et al. (2019) discussed the  
288 performance of secondary wastewater treatments for the removal CECs that can be implicated  
289 in wastewater reuse practices.

290

291 **3. Methodological approach to derive chemical quality standards for RWAI**

292 **3.1 Identification of the environmental compartments, populations and individuals at risk**  
293 **of direct or indirect exposure**

294 The EC reclaimed water reuse regulation proposal (EU, 2018) requires that MS ensure that the  
295 use of RWAI has no adverse effects on environmental matrices: soil, groundwater, surface  
296 water, and dependent ecosystems, including crops to be irrigated. Indeed, both humans and  
297 other organisms in the environment can be exposed directly (i.e. receipt of irrigation water,  
298 splashes or spray drift) or indirectly (i.e. bioaccumulation within the food web) during water  
299 reclamation and agricultural irrigation operations. However, the direct exposure of workers and  
300 residents in the framework of operational safety is not addressed in this paper that focuses solely  
301 on environmental assessment (including humans via the environment).

302

303 Under the environmental assessment, both the soil and the water compartments can be exposed  
304 to reclaimed water releases at different levels according to the type of crops and farming  
305 practices adopted. The relevant soil compartment is the agricultural landscape and the pertinent  
306 water compartments are water bodies in the vicinity of the agricultural landscapes that could  
307 receive water releases during irrigation operation. Depending on the location of the agricultural  
308 landscapes, these water bodies include surface waters (i.e. rivers, lakes), coastal zones (i.e.  
309 marine water) and groundwater. As a consequence, the organisms to be protected (defined as  
310 protection targets) are those living in these ecosystems, directly exposed to reclaimed water but  
311 also the organisms that consume water and/or food (predatory organisms) from these  
312 ecosystems and can hence be indirectly exposed through the accumulation of the chemical  
313 contaminants in the food chain (secondary poisoning). The predatory organisms are defined as  
314 predators or top predators, depending if a simple or more complex food web is assumed.

315 For water bodies, the populations at risk considered in several EU regulations are pelagic  
316 organisms, benthic organisms, predators and top predators (for the marine food web) and the

317 individuals at risk are humans via ingestion of drinking water and fishery products. For the soil  
318 compartment, the populations considered as at risk are the terrestrial communities  
319 (microorganism, invertebrates, plants), worm-eating predators (birds and mammals) and top  
320 predators of these ecosystems, and the individuals at risk are humans consuming animal  
321 products (meat, milk, eggs) and crops.

322

### 323 **3.2 Evaluation of the applicability of the EQS derivation methodology under the WFD to** 324 **the use of RWAI**

325 The German Environment Agency (UBA, 2017) emphasised that quality requirements for  
326 reclaimed water reuse should comply with and complement the current EU legislation that  
327 already exists for surface water and groundwater protection, and in particular the WFD  
328 principles and resulting quality standards. The EC proposal for a water reuse regulation (EU,  
329 2018) also states that, amongst others, the regulatory requirements of the WFD and its daughter  
330 EQS and PS Directives have to be fulfilled. Therefore, we believe that the relevance of using  
331 EQS for RWAI and the potential need for further adaptation must be investigated.

332 In 2005, a technical guidance document was prepared (Lepper, 2005) for the purpose of EQS  
333 derivation that comply with the requirements of Annex V of the WFD. It was further updated  
334 in 2011 to develop the steps required to derive EQS for metals, and EQS for biota and sediment  
335 (TGD, 2011). Recent developments specify the methodology for the use in the derivation of  
336 biota standards which address human health and secondary poisoning of wildlife, and the  
337 derivation of EQS for bioavailable metals (TGD, 2018). It is important to note, however, that  
338 the EQS are defined in relation to the protection of organisms from direct chemical exposure in  
339 surface waters and organisms from indirect exposure via the aquatic food chain (secondary  
340 poisoning) and that the use of RWAI is out of scope of EQS derivation. In order to assess if  
341 existing EQSs are relevant for this new purpose, the protection targets of the EQSs have been

342 compared with those identified for RWAI. Practically, the methodology for EQS derivation  
 343 involves the previous derivation of quality standards (QSs) for each identified protection target  
 344 with a Maximum Acceptable Concentration (MAC) and an Annual Average (AA)  
 345 concentration established in surface waters and AA concentrations established in biota and  
 346 sediment. The protection targets of RWAI are listed in the Table 2 together with their relevant  
 347 identified applicable QS from EQS derivation methodology. The absence of existing QS for  
 348 some protection targets is also indicated.

349

350 **Table 2: Protection targets to be addressed for the use of RWAI and existing relevant**  
 351 **Quality Standards from the TGD reaching this goal**

Protection target		Related exposure source	Relevant QS of the TGD
Freshwater ecosystem	Pelagic organisms	Freshwater	MAC-QS <sub>fw, eco</sub> AA-QS <sub>fw, eco</sub> (µg.L-1)
	Benthic organisms	Freshwater sediment	QS <sub>sediment, fw</sub> (µg.kg-1sed ww or dw)
	Predators (birds and mammals)	Freshwater prey (e.g. fish and molluscs)	QS <sub>biota fw, sec pois</sub> (µg.kg-1biota)
Marine ecosystem	Pelagic organisms	Saltwater	MAC-QS <sub>sw, eco</sub> AA-QS <sub>sw, eco</sub> (µg.L-1)
	Benthic organisms	Saltwater sediment	QS <sub>sediment, sw</sub>



			( $\mu\text{g.kg}^{-1}\text{sed ww or dw}$ )
	Predators (birds and mammals)	Saltwater prey (e.g. fish and molluscs)	$QS_{\text{biota sw, sec pois}}$ ( $\mu\text{g.kg}^{-1}\text{biota}$ )
	Top predators (e.g. killer whales and polar bears)	Birds and mammals	$QS_{\text{biota sw, sec pois}}$ ( $\mu\text{g.kg}^{-1}\text{biota}$ )
Terrestrial ecosystem	Agricultural soil organisms incl. crops	Agricultural soil	No existing QS
	Predators (birds and mammals)	Terrestrial prey (e.g. earthworms and plants incl. crops)	$QS_{\text{biota fw, sec pois}}$ ( $\mu\text{g.kg}^{-1}\text{biota}$ )
	Top predators (e.g. raptors and mustelids)	Small birds and mammals	No existing QS
Humans	Humans consuming fishery products	Freshwater and saltwater fish, molluscs...	$QS_{\text{biota, hh food}}$ ( $\mu\text{g.kg}^{-1}\text{biota}$ )
	Humans consuming drinking water	Abstracted water and Groundwater	$QS_{\text{dw, hh}}$ ( $\mu\text{g.L}^{-1}$ )
	Humans consuming irrigated crops and animal products	Crops Meat, milk and eggs	No existing QS

352

353 3.2.1 Protection targets covered by existing QSs

354 As expected, Qs derived to protect aquatic organisms, their predators/ top predators and  
355 humans during drinking water and fishery products consumption under the WFD Common  
356 Implementation Strategy will also be protective for the use of RWAI. More surprising, the QS  
357 for biota protecting predators consuming freshwater fish ( $QS_{\text{biota fw, sec pois}}$ ) and derived according  
358 to the diet-based concentration method of the TGD (2011) could also be considered protective  
359 for terrestrial predators like birds and mammals eating worms and/ or plants because the  
360 toxicological data on birds and mammals can be considered also relevant for terrestrial  
361 predators. However, this methodology makes no difference between various food items and use  
362 a default factor of 3 to correct for the differences in caloric content between standard laboratory  
363 food on the one hand and prey species in the field on the other hand. According to Verbruggen  
364 (2014) this default factor is reasonable for fish (factor 2.8) but is not for earthworms (factor  
365 5.2), which have a much lower caloric content based on fresh weight. It is probably also the  
366 case for plants meaning that  $QS_{\text{biota fw, sec pois}}$  could be insufficiently protective for terrestrial  
367 worm-eating and plant-eating predators.

### 368 *3.2.2 Non-covered protection targets by existing Qs*

369 Other organisms exposed via the soil (by direct or indirect exposure to reclaimed water) are not  
370 covered by an existing QS:

- 371 - The terrestrial organisms including microorganisms, invertebrates, plants via direct  
372 exposure and top predators consuming small birds and mammals via indirect exposure  
373 through bioaccumulation in the food chain.
- 374 - Humans via indirect exposure through bioaccumulation in the food chain (crops and  
375 animal products).

376

### 377 **3.3 Proposal for the adaptation of existing QS or the development of new QS for RWAI**

378 **use:**

379 For pelagic organisms (and benthic organisms if the Equilibrium Partitioning Method is used),  
380 the Qs of the WFD can be considered over-protective because reclaimed water will be diluted  
381 within receiving water bodies. Increasing the Qs concentration by the factor of dilution of  
382 receiving freshwater and marine water bodies as normally considered in chemical risk  
383 assessment could be implemented for RWAI.

384 For terrestrial organisms including microorganisms, invertebrates and plants, the methodology  
385 to derive Predicated Non-Effect Concentration for the soil (PNEC<sub>soil</sub>), in different EU  
386 regulatory frames is recommended to develop a new Q<sub>soil</sub> (ECHA, 2008). However, it should  
387 be noted that terrestrial data are available only rarely for many chemicals and the equilibrium  
388 partitioning method commonly used in case of terrestrial data lacking is driving a high  
389 uncertainty on the estimated PNEC<sub>soil</sub>. Moreover, the actual scoping of PNEC<sub>soil</sub> derivation  
390 does not include terrestrial invertebrates living above-ground (e.g. ground dwelling beetles),  
391 terrestrial vertebrates living a part of their lifetime in soils (e.g. mice) and groundwater  
392 organism (invertebrates and micro-organism).

393 For terrestrial predators (birds and mammals) and top predators like raptors and mustelids  
394 consuming small birds and mammals, the caloric content-based diet concentration methodology  
395 described by Verbruggen (2014) can be adopted when both the energy content and  
396 bioaccumulation parameters are available for several food items in order to select the critical  
397 food item in the food chain that is most relevant for secondary poisoning in agricultural soils.

398 For humans exposed through the terrestrial food chain, the existing general food standards  
399 established for crops and animal products for relevant population groups could be used as a first  
400 instance. For example, the WHO Codex Alimentarius (including food additives, residues of  
401 pesticides, veterinary drugs and contaminants), the EC Maximum Residues Levels for  
402 pesticides in food (EU, 2005) and the EC maximum levels for certain contaminants in  
403 foodstuffs defined for selected metals, PCBs, dioxins and polycyclic aromatic hydrocarbons

404 (EC, 2006) could be drawn upon. If such food standards are not available for the substances  
405 and/or food of concern, the methodology used to derive the QS intended to protect humans  
406 against adverse health effects from consuming contaminated fishery products ( $QS_{\text{biota, hh food}}$ ) in  
407 the TGD (2018) could be adapted considering all the various sources of food consumption of  
408 humans (fish, crops and animal products).

409

#### 410 **4. Data gaps, uncertainties and research needs**

##### 411 **4.1 Chemical contaminants in reclaimed water: transformation products and unregulated** 412 **substances**

413 Urban WWTPs and water reclamation plants are designed to remove chemical substances via  
414 various biotic and abiotic degradation processes. Therefore, transformation products, also  
415 referred to as metabolites, are expected to be generated at treatment plant facilities. For  
416 example, Paredes et al. (2019) demonstrated that in a secondary wastewater effluent resulting  
417 from photocatalytic degradation, 156 transformation products originating from eight  
418 pharmaceuticals could be detected. However, in most of the studies designed to assess  
419 wastewater treatment removal efficiency, transformation products are not identified as the main  
420 focus is on the disappearance of the parent substance. It is an important issue since it has been  
421 reported that transformation products can sometimes represent a higher toxicity concern than  
422 parent substances. For example, carbamazepine-10,11-epoxide, a main metabolite of the  
423 antiepileptic drug carbamazepine, is reported to exhibit a higher chronic toxicity on the midge  
424 *Chironomus riparius* in comparison to the parent compound (Heye et al., 2016). Therefore, it  
425 is highly recommended that future investigations include the identification and quantification  
426 of transformation products and assess their toxicity in complementary studies. Another  
427 uncertainty is related to the occurrence of unregulated “substances” including those transported  
428 in storm water runoff, illicit substances etc., that could reach municipal WWTPs as, for

429 example, reported in the Netherlands and Spain by Bijlsma et al. (2012 and 2014) and, more  
430 recently, microplastics that are not removed by WWTP (Lares et al., 2018) and persist in the  
431 environment (De Souza Machado et al., 2018). Therefore, a better characterisation of the impact  
432 of storm water run-off and microplastics, as well as the occurrence of illicit substances in urban  
433 WWTPs, is required.

434

#### 435 **4.2 Prioritisation and QS derivation methodology: Data on terrestrial ecosystems and QS** 436 **biota conversion**

437 Soil is a primary exposed environmental compartment within RWAI and, whilst exposure after  
438 sludge application is considered within regulatory chemical risk assessment, the use of RWAI  
439 is not. Data recorded over many years of RWAI practice in the Tula Valley (Mexico) show a  
440 mean retention of 86% of chemical contaminants in soil (Navarro et al., 2015) and Carter et al.  
441 (2019) recently reported that RWAI continuous application has resulted in pharmaceuticals  
442 building up in soils to total concentrations of up to ca. 15 mg kg<sup>-1</sup>. However, regulatory data on  
443 soil are mostly missing. Indeed, until recently, data on freshwater aquatic ecosystems were  
444 considered the minimum necessary dataset under most regulations. Biocides and PPPs  
445 regulations are an exception, requiring the submission of terrestrial data on corresponding  
446 active substances. In order to perform a targeted prioritisation exercise of contaminants in  
447 RWAI and derive sound QS for the soil communities, it is necessary to investigate their  
448 terrestrial persistence, bioaccumulation through the food chain and toxicity including endocrine  
449 disrupting effects.

450 Finally, the EC methodology to derive EQS supports the conversion of QSs developed for biota  
451 in the aquatic food web into equivalent aquatic concentrations in order to allow monitoring in  
452 water only. For the conversion of a terrestrial biota standard into a soil concentration, a similar  
453 calculation can be made using the relevant biota-to-soil-accumulation-factor (BSAF, usually

454 for earthworms) and biomagnification factors from preys to predators (usually from earthworm  
455 to small terrestrial birds or mammals) as suggested by Verbruggen (2014). In order to use a  
456 similar methodology for humans consuming irrigated crops and animal products, it would be  
457 necessary to develop BSAFs in the different irrigated crops and bio transfer factors (BTF) in  
458 animal products (e.g. milk, meat, eggs). Finally, a conversion of soil concentration into water  
459 concentration would require the development of models which estimate the soil concentration  
460 resulting from the repeated application of RWAI to fields, as currently undertaken for sludge  
461 applications in most of the chemical regulations (e.g. REACH, BPR, medicinal products for  
462 human use).

463

## 464 **5. Conclusions**

465 A barrier to the greater use of RWAI is the perceived risks the practice may pose to human  
466 health and the environment, especially related to CECs identified as not fully removed within  
467 urban WWTPs or reclamation plants. This analysis has shown that current legislation pertaining  
468 to industrial releases, WWTPs and environmental protection is not sufficient to manage the risk  
469 posed by hazardous substances within a reclaimed water reuse context. This is understood to  
470 be a function of the fact that existing legislations were not specifically developed to address  
471 reclaimed water reuse pathways and protection targets. Thereby, almost all chemicals  
472 categories intentionally manufactured or the by-products of industrial processes could reach  
473 urban WWTP. The vast majority of these chemicals being regulated under the authority of the  
474 European agencies, it is therefore highly recommended to use the output of their prioritisation  
475 methods, such as the SVHC roadmap by ECHA, in order to identify and prioritise the  
476 substances of concern for RWAI. Special attention, however, should also be given to excluded  
477 chemical categories (i.e. substances out of scope from a European agency perspective) when  
478 undertaking the specific prioritisation exercise together with consideration of the relevance of

479 prioritisation criteria for the reuse of RWAI. The declassification of substances from the priority  
480 list should be made possible based on data on removal efficiencies of wastewater treatment  
481 technologies. Once a list of priority substances has been defined, QS can be derived in order to  
482 monitor the chemical quality of reclaimed water and ensure compliance with human health and  
483 environmental protection goals. Under the EU WFD, a methodology is available to derive EQS  
484 for contaminants in order to protect water bodies from chemical pollution. In the present  
485 analysis we conclude that while the same methodology is appropriate with regard to protection  
486 of the water ecosystems (and their consumers) from run-off during irrigation, the main exposed  
487 organisms in soil or via the terrestrial food chain would not be adequately protected under this  
488 approach. In parallel with ongoing research, further work will draw on knowledge already  
489 available under different regulatory frameworks to develop such a methodology including, for  
490 example, opportunities to adapt current QSs or derive new QSs for terrestrial protection targets  
491 as required. Overall, this work provides guidance to policy makers, researchers and  
492 practitioners with regard to both the need and opportunities for the derivation of QS for CECs  
493 in RWAI. Key data gaps have been identified concerning soil ecosystems, wastewater treatment  
494 technology removal efficiencies and the occurrence, behaviour and fate of transformation  
495 products generated during the reclaimed water reuse cycle requiring further investigations at  
496 both research and regulatory levels. As a future perspective, we believe that in order to facilitate  
497 moves towards managing the risks posed by chemicals in the context of RWAI, standardised  
498 emission scenarios should be developed for this use, supporting systematic assessment of risks  
499 posed by regulated chemical substances released from WWTPs.

500

#### 501 **Conflicts of interest**

502 The authors have no conflict of interest to declare regarding this article.

503

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512



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