1	Kleptoparasitism in gulls (<i>Laridae</i>) at an urban and a coastal foraging
2	environment: An assessment of ecological predictors.
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13	Key words: Kleptoparasitism, urban ecology, coastal ecology, Laridae.

14 **ABSTRACT**

- 15 **Capsule:** Kleptoparasitism in gulls occurred at a greater rate at an urban compared with
- a coastal site. Population density and prey size predicted the rate of kleptoparasitism atthe urban site.
- 18 **Aims:** To investigate and assess the ecological variables associated with
- 19 kleptoparasitism among gulls at urban and rural sites.
- 20 Methods: Field observations were conducted at Brancaster (coastal rural) and
- 21 Billingsgate Market (urban) to examine differences in the rate of kleptoparasitism in
- 22 mixed-species flocks of gulls. Four key variables (prey size, population density, season
- and species) were assessed as predictors of kleptoparasitism.
- 24 **Results:** Generalized linear models revealed significant effects on kleptoparasitism rate
- of site, population density and prey size, and two-way interactions between these main
- terms. Population density and prey size differed significantly between sites, but
- 27 population density appeared to predict the rate of kleptoparasitism.
- 28 **Conclusion:** Kleptoparasitism may well aid invasion and increase the range of
- 29 environments a gull can tolerate by helping them meet their energy needs in novel
- 30 environments where normal foraging behaviours are difficult to implement.
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34 INTRODUCTION

35 Kleptoparasitism can be defined most simply as "seizing food gathered by another" 36 (Hatch 1970) and it is a foraging strategy that can reduce the costs of searching for and 37 procurement of food (Thompson 1986, Barnard 1990, Giraldeau & Caraco 2000, Giraldeau & Dubois 2008). Kleptoparasitism has been observed within and between 38 39 species across various animal taxa (Brockmann & Barnard 1979) including insects (Erlandsson 1988), spiders (Coyle et al. 1991), mammals (Höner et al. 2002), and even 40 41 humans (Schoe et al. 2009). Birds, more than other taxa, have been the focus of 42 detailed research in this area (Barnard 1990). In a seminal and exhaustive review of the 43 avian literature, Brockmann & Barnard (1979) identified gulls Laridae as the most prevalent kleptoparasites and listed ecological factors conducive to kleptoparasitism. 44 45 These included large concentrations of hosts (victims) and quantities of food, and large, high quality food items to steal. Here, we focus on kleptoparasitism within and between 46 47 gull species in two distinct ecologies.

Research on gulls has demonstrated that kleptoparasitism is a facultative behavioural strategy that is applied flexibly as ecological conditions vary. Field research conducted by Maniscalco & Ostrand (1997) found that the degree of kleptoparasitism covaries with the concentration of hosts. In mixed-species foraging flocks of seabirds the rate of kleptoparasitism committed by gulls increased as the density of birds increased. The distribution of food resources covaried with foraging flock density, such that food distribution should directly predict rates of kleptoparasitism.

55 Hamilton (2002) explored the relationship between food availability and 56 kleptoparasitism by assessing the distribution of foragers between patches that were 57 high or low in the rate of food input. The model showed that less kleptoparasitism occurred in high resource input patches and that kleptoparasitism increased with 58 59 decreasing resource input. Increasing the density of competitors in a patch had the effect of increasing the proportion of kleptoparasites, in line with the observations of 60 61 Maniscalco & Ostrand (1997). Increasing the number of competitors and decreasing 62 resources increased the intensity of resource competition, thus promoting kleptoparasitism. 63

The findings of Maniscalco & Ostrand (1997) and Hamilton (2002) strongly indicate that kleptoparasitism is a flexible behavioural strategy. However, no research has yet investigated the value of kleptoparasitism in evolutionarily novel environments. The increasing urbanisation of gull populations provides an opportunity to address this question. Many gull species that are found in urban settings are of conservation concern due to declining populations over the last 25 years or more (Eaton *et al.* 2015). Understanding the behavioural strategies adopted by these species will help to further
understand the stresses they face (Eaton *et al.* 2015, Reid 2004, Madden & Newton
2004).

73 Gulls are found in a variety of habitats (Pons et al. 2005) and have been noted 74 for their tendency to colonise novel environments (Rock 2005). As members of the sub-75 order Charadrii they are part of only two groups of modern birds known to have been present since the Cretaceous period over 65 million years ago (Proctor & Lynch 1993). 76 77 Their significant investment in kleptoparasitism highlights the importance of this strategy 78 to gull species, and suggests kleptoparasitic behaviours may have played an important 79 role in aiding the radiation and longevity of this family. For any animal, invading an 80 urban environment entails finding solutions to a number of survival problems, the most 81 pressing of which is acquiring food. Kleptoparasitism is a foraging strategy that can be used by gulls when very few conditions are met. The only essential conditions are the 82 83 availability of opportunistic food sources and the presence of other foragers, of any species, to provide cues as to the location of food. 84

In this study, we investigated the predictors of the rate of kleptoparasitism in 85 foraging gull populations across two environments with contrasting parameters: 1) a 86 87 coastal environment, assumed to be typical of the shoreline foraging ecologies 88 encountered by gulls throughout much of their evolutionary history and 2) an urban 89 environment. Our predictions were: 1) that population density would be positively 90 related to the rate of kleptoparasitism, 2) that large prey items would be more 91 susceptible to theft, so that prey size should covary with kleptoparasitism rate, 3) that 92 kleptoparasitism would vary between breeding (March to August) and non-breeding 93 seasons, 4) that gull species would differ in their use of kleptoparasitic behaviour and 5) 94 that there would be a higher rate of kleptoparasitism at the urban site, where food was 95 predicted to be a rapidly decreasing resource.

96

98 METHODS

99 Study Sites

100 Research was conducted at two sites in the UK, between July 2014 and June 2015. Three 101 days of observations were conducted at each site in all calendar months. Site 1 was Brancaster Beach, Norfolk, UK (52°58'30"N 0°38'11"E); a public beach situated within a 102 saltmarsh environment, managed by the National Trust. The low tide at Brancaster 103 104 exposes a large area of sand beach where numerous bird species forage for prey items exposed by the receding tide. The study area was a section of beach demarcated by two 105 106 water channels where the outflow of water draining from the saltmarsh joins the sea. 107 These two channels marked the eastern and western boundaries of the study area. At its maximum, when the tide was at its lowest point, the study area covered 1.6 km². Site 2 108 was an urban location at Billingsgate Market, east London (51°30'20"N 0°00'43"W); a 109 110 commercial fish market. Research was conducted in a car park used by fish merchants to process and load stock onto vans, and covered an area of 0.0104 km². The size of study 111 112 areas was calculated using scaled aerial photographs from Google Maps.

113

114 Study species

Large numbers of gulls aggregated daily to forage at both sites. At Brancaster, gulls 115 aggregated to forage in mixed-species groups over a period of about four hours; two 116 hours either side of the low tide mark. The most numerous gull species present at this 117 118 site were Great Black-backed Gull Larus marinus, Herring Gull Larus argentatus, Blackheaded Gull Chroicocephalus ridibundus and Common Gull Larus canus. These four 119 120 species were the focus of research at Brancaster throughout the year. However, the numbers of Common Gulls and Black-headed Gulls were lower during the summer 121 122 months as birds were mostly away breeding elsewhere. The foraging population of gulls at Brancaster had a mean daily population size of 176 (Range: 0 - 924) gulls. The daily 123 124 mean number (and % of population) of each species at the site was: Herring Gull 75 125 (42%), Black-headed Gull 64 (37%), Common Gull 31 (18%) and Great Black-backed 126 Gull 6 (3%).

Gulls aggregated at Billingsgate to exploit seafood waste and food leftovers
discarded in the car park areas. The population consisted of Great Black-backed, Herring
and Black-headed Gulls, and all three species were the focus of research. Great Blackbacked and Herring Gulls were present at Billingsgate all year round, with some birds
nesting on the surrounding market buildings. Black-headed Gulls were largely absent
from the site between the April and July when they were away breeding elsewhere. The

foraging population at Billingsgate had a mean daily population size of 40 (Range: 24 79) gulls. The daily mean number (and % of population) of each species at Billingsgate
was: Herring Gull 28 (70%), Black-headed Gull 7 (19%) and Great Black-backed Gull 4
(11%).

137

138 Measuring kleptoparasitism

139 Population size and composition were calculated using scan samples at 30 minute 140 intervals. All attempts at kleptoparasitism, both successful and unsuccessful, were counted. Kleptoparasitic behaviours included the use of force to take items from another 141 142 bird, theft without any interaction with the host and simultaneous theft (or attempted 143 theft) from a host, by two or more competitors (Giraldeau & Caraco 2000). Size of food 144 items was estimated visually in relation to bill lengths. The Herring Gull bill length was 145 used as a standard comparative measurement to assess the length of food items (Table 1). Food items contested by the smaller or larger species of gulls were calibrated in 146 147 relation to the mean size of Herring Gull bills. Prey sizes were assigned to increasingly 148 large categories in relation to Herring Gull bill size (0.5, 0.75, 1, 2, 3, 4, 5 or 6).

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- 150

[INSERT TABLE 1 ABOUT HERE]

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152 Field observations

Brancaster: The majority of field sessions were conducted at low tide to coincide with the
presence of the largest numbers of birds. Field sessions had a mean duration of 120
minutes (range: 90 - 220 minutes) and a total of 74 hours 28 minutes of observations
were conducted at Brancaster over 34 days. Field sessions were conducted during
daylight between 04:45 hours and 17:00 hours, to permit behaviour to be recorded on
video.

159 An observer (RS) scanned the beach for gulls, using binoculars, and then 160 approaching to within a distance from which foraging patches could be identified, the gulls counted and their foraging behaviour recorded on video. Patches were classified as 161 162 discrete clusters of two or more birds exploiting a food source or searching for food in a given area. The location of patches was constantly shifting around the beach. Once a 163 164 patch was identified, video recording was carried out from a fixed position with a Sony 165 8.9 megapixel HD camcorder mounted on a tripod. The distance from which a patch was 166 filmed varied and was contingent on how closely the researcher could approach without

disrupting the foraging birds. Recording of a patch was concluded when the gulls in that
patch dispersed. The duration of recordings ranged from 32 seconds to 40 minutes 47
seconds, and 142 patches were recorded in total. Videos were analysed at a later date
for incidents of kleptoparasitism.

Gull counts were conducted every 30 minutes from the start of a field session, using a series of digital images (Nikon Coolpix P510 42x zoom) taken to record all gulls in the study area. These images were analysed at a later date to calculate the population size and composition.

Billingsgate: Observations at Billingsgate were conducted between 07:00 hours and
15:00 hours. The trading hours for the market were 03:00 hours to 08:30 hours. After
08:30 hours, the main activity at the site is the clean-up of the market and car park.
Prior to 07:00 hours it was not possible to conduct observations due to the large number
of vehicles at the site. Field sessions were conducted on days when the market was
operational (Tuesday to Saturday) and when it was closed (Sunday and Monday).

Observations and recordings were conducted from a vehicle placed to give the best vantage point on the day and the position varied between observation days. Gull activity at foraging patches was video recorded by the observer (RS) until the resource in the patch had depleted and the birds dispersed. Over the study, 183 foraging patches were recorded and these ranged in duration from 25 seconds to 29 minutes 36 seconds. For each recording we noted the number of each species and ages of the gulls present. Patch videos were analysed at a later date for incidents of kleptoparasitism.

Gull counts were conducted at 30 minute intervals from the start of a field session. As at Brancaster, a series of digital images were taken of the study area, and later analysed to calculate the population at the time of the sample, including the species and ages of gulls present. The same recording and photographic equipment used at Brancaster was used at Billingsgate.

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194 Statistical analysis

All statistical analyses were conducted using R version 3.3.1 (R Core Team, 2016).

196 Figures were produced in R using the base package or ggplot2 version 2.1.0 (Wickham,

197 2009).

198 **RESULTS**

In total, there were 595 kleptoparasitic events recorded at Billingsgate and 99 at
Brancaster. A single event involving a Lesser Black-backed Gull *Larus fuscus* was
excluded from the analysis. There were 15 missing data points for prey size and all data
associated with these observations were also removed.

203 Data were analysed using a series of generalized linear models (GLM). The 204 response variable was the rate of kleptoparasitism (events per minute) for each patch 205 observed. This variable was \log_n transformed to improve the dispersion of data. The 206 predictor variables were population density (daily mean population density (birds.km⁻²) 207 over each study area), prey size, season (breeding/non-breeding), species of 208 kleptoparasite and site (Billingsgate/Brancaster). The population density, prey size and response variables were each centred and scaled in order to standardize them. 209 210 Standardizing continuous variables facilitated meaningful comparisons between model 211 coefficients. All variables were entered as main effects into a GLM using a Gaussian 212 function with identity link (Table 2).

To avoid over-fitting a relatively small dataset, we next adopted a subtractive 213 214 approach using a backward step() function. This is a sequential approach to model fitting using the Akaike Information Criterion corrected for finite sample sizes (AICc). 215 216 Models with AICc values which differ by less than 2 are regarded as being similarly good fits, whereas differences (Δ AICc) over 2 suggest the model with the higher AICc is a 217 218 poorer fit (Burnham & Anderson 2001). This approach led to the removal of season and 219 species from the best fitting model, leaving a minimal main effects model with site, 220 standardized population density and standardized prey size as significant predictors of 221 kleptoparasitism rate (Table 2).

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223

[INSERT FIGURES 1 & 2 AND TABLE 2 ABOUT HERE]

224

A positive correlation was apparent between population density and standardized log_n
kleptoparasitism rate at Billingsgate, and to a lesser extent at Brancaster (Figure 1). A
positive correlation was also seen between prey size and standardized log_n
kleptoparasitism rate at Billingsgate, but this correlation appears to be negative at
Brancaster (Figure 2).

Owing to the trend differences noted between sites (Figures 1 and 2), a second(Gaussian) GLM was constructed which included the main effects and all pairwise

interactions between site, standardized prey size and standardized population density.
This model suggested that the interactions were significant (Table 2). Notably, however,
the coefficients for the main effects of site, standardized prey size and standardized
population density remained approximately the same (Table 2) indicating that the data
collected at Brancaster did not evidence a trend reversal.

Data from Brancaster showed a smaller range of values for prey size and population density than Billingsgate, and in the case of population density the data ranges for the two sites were non-overlapping (x axes in Figures 1 and 2, Table 3). This suggests that additional data may be needed to be collected at Brancaster (or an equivalent site) to capture the influence of these variables on kleptoparasitism across this reduced ranges. Prey sizes and population densities were both significantly greater at Billingsate (Table 3).

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- 245

[INSERT TABLE 3 ABOUT HERE]

Returning to the response variable and to explore the relative importance of pairwise 246 247 interactions in the interactions model we again adopted a subtractive Δ AICc approach (using the drop1() function). All three pairwise interactions between the three entered 248 variables could be removed while lowering the AICc with one exception: the site × prey 249 size interaction. Removing only this interaction from the model resulted in an increased 250 251 AICc. The reasonable fit offered by models of kleptoparasitism that include interactions and the significant differences between the two key continuous predictor variables 252 253 support the view that the two sites show a core ecological difference.

254 **DISCUSSION**

Our results show that rates of kleptoparasitism in gulls are affected by bird population density and prey size. While these two predictors seem to be robustly positively correlated with increased rates of kleptoparasitisim, differences between the sites in the predictors and the response variable support a more cautious interpretation. It is possible that the population density measure may under-estimate the search time entailed for would-be kleptoparasites at Brancaster, in contrast to Billingsgate.

261 Maniscalco & Ostrand (1997) demonstrated how a shift in the foraging conditions 262 in a single environment resulted in an increase in kleptoparasitisic behaviour by gulls. 263 Our results complement that finding and describe clear frequency differences between a 264 rural and an urban site. Our study gave focus to the role that kleptoparasitism might 265 play in helping gulls meet their energy needs when invading urban environments. Our 266 sample size is small – one rural and one urban environment only – and so we must be cautious in generalizing any findings. However, here we consider the mechanism by 267 268 which urban invasion might occur when birds initially encounter those spaces. As Marzluff et al. (2001) have described, the foraging environment changes on a gradient 269 270 between wildland environments and heavily urbanised spaces. As a gull moves along this 271 gradient it will encounter fewer aspects of the foraging environment with which it is 272 familiar. At one extreme of this gradient are heavily urbanised spaces containing lots of 273 anthropogenic novelty with few of the characteristics of a natural food web. Billingsgate 274 is such a site, as all the food there is waste provisioned by human activities. A gull, or 275 small founder population of gulls, invading an urban environment must either obtain 276 food through: innovation by exploring the novel foraging space; learning what things are 277 edible and where to find them by copying other foragers; or stealing food from other 278 individuals.

279 Previous research has emphasised the importance of innovative behaviour in 280 species that invade novel environments (Sol et al. 2005, Sol et al. 2008). Some 281 proportion of innovators, who explore the environment, are essential in any population 282 otherwise there would be no one to copy, however, rates of innovation in any natural population are quite low (Page 2008), making the second two options, copying or 283 284 stealing, more likely for invasive gulls. Of these, copying other foragers to learn the locations of food is necessary, but stealing can be efficient as it avoids the cost of the full 285 286 foraging cycle by acquiring food discovered by another's effort. To this end, we can think of stealing information and stealing food as on a continuum; the cost of stealing the food 287 288 is actually the marginal cost of stealing conditional on already observing the victim.

As gulls can range over large areas they can occupy different habitats on the gradient described above in a single foraging bout or day. This suggests a plausible 291 alternative mechanism to the one described above by which gulls can invade urban 292 spaces. They can forage in environments where they encounter conditions with which 293 they are familiar, and then make forays and encroachments into urban spaces in search 294 of supplemental food and other foraging populations. Foraging in the littoral zone of 295 rivers and estuaries that pass near or through cities is an example of how this might 296 work. In our argument regarding the utility of kleptoparasitism we have made the 297 assumption that while this second mechanism would undoubtedly occur, the additional 298 travel and foraging costs involved would make it energetically more expensive than 299 exploiting other foragers through kleptoparasitism. These costs would not be prohibitive 300 of this foraging approach but we suggest that these additional costs would make it 301 secondary to the kleptoparasitic strategies considered above. At present little is known 302 about the foraging habits and ranges of urban gull populations, and tracking of 303 individuals to see the extent of their foraging ranges was not undertaken in our study. 304 Further work of this nature would be useful to assess whether urban gull populations, 305 such as the one at Billingsgate, obtain their food exclusively at that site or exploit a 306 variety of habitats.

307 Comparing the range and diversity of environments occupied by different gull 308 species may also provide insights into the role of kleptoparasitism in supporting invasion. 309 As described by Brockmann & Barnard (1979) 23 of 88 gull species are reported to use 310 kleptoparasitism. A comparison between kleptoparasitic and non-kleptoparasitic gull 311 species of the diversity of habitats that these species routinely breed and forage in, and 312 the extent of their ranges, may provide indirect evidence for the proposed role of 313 kleptoparasitism as a strategy that buffers against environmental change. If this 314 proposed function of kleptoparasitism is plausible we should expect to see kleptoparasitic 315 gull species occupying a broader diversity of habitats than non-kleptoparasitic gull 316 species.

317

318 Ecological predictors of kleptoparasitism

Birds are dispersed widely at Brancaster and population density was positively correlated with kleptoparasitism. This finding is consistent with empirical research that shows foragers are more likely to encounter each other at high population densities and interact aggressively (Colwell 2010). It also lends support to the theoretical model of Hamilton (2002) demonstrating increased kleptoparasitism with increasing competitor density.

In addition, King *et al.* (2009), in research on baboon troops, found that the distribution of resources influenced the rate of kleptoparasitism, with small tightly 327 clustered patches prompting more kleptoparasitism. The resources in the study 328 environments at Brancaster and Billingsgate were clearly distributed in different ways 329 that provided a useful test of the pattern of behaviour described by King et al. (2009). 330 The food at Brancaster was naturally more widely dispersed throughout the environment and it was a much larger study area than Billingsgate, permitting gulls to be more 331 spaced out whilst foraging. Our results support the finding of King et al. (2009) with 332 333 more kleptoparasitism in high density areas where gulls were forced to forage closer 334 together, perhaps making it easier to take advantage of opportunities to try and steal.

335 The finding that larger food items were more likely to be targeted for theft is in 336 accord with the findings of several empirical studies (Hopkins & Wiley 1972, Ens & Cayford 1996, Leeman et al. 2001). Large prey items provide a conspicuous visual cue 337 338 to the presence of food, and a large prey item requires longer handling time, which increases the likelihood that other foragers can approach and try to steal the item. Our 339 340 use of an ordinal scale to measure the size of prey items in terms of bill lengths was a limitation of this study that may have resulted in a loss of additional data of value. 341 342 Ordinal measures of this kind result in a lack of precision that may mask useful patterns 343 of behaviour. A more finely grained measure would have been beneficial, but the ordinal measure of bill lengths used represented the most practical way to assess the size of 344 345 food items from the observational data obtained.

346 *Concluding remarks:* Kleptoparasitism was higher in the urban environment. Large prey 347 items that take longer to handle, and high population densities that increase competition 348 for available resources were the critical aspects of the urban environment at Billingsgate 349 that promoted kleptoparasitism. Kleptoparasitism may well aid invasion and increase the 350 range of environments a gull can tolerate by helping them meet their energy needs in 351 novel environments where normal foraging behaviours are difficult to implement.

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357 References

- Barnard, C.J. 1990. Parasitic relationships. In C.J. Barnard & J.M. Behnke (eds.). *Parasitism and Host Behaviour* pp. 1-33. Taylor & Francis, London.
- Brockmann, H.J. & Barnard, C.J. 1979. Kleptoparasitism in birds. *Animal Behaviour*27: 487-514.
- Burnham, K.P. & Anderson, D.R. 2001. Kullback-Leibler information as a basis for
 strong inference in ecological studies. *Wildlife Research* 28: 111–119.
- 364 Cohen, J.E., Jonsson, T. & Carpenter, S.R. 2003. Ecological community description
 365 using the food web, species abundance, and body size. *Proceedings of the National* 366 Academy of Sciences 100: 1781-1786.
- 367 Colwell, M.A. 2010. Shorebird Ecology, Conservation, and Management. University of
- 368 California Press, Berkeley, USA.
- Coyle, F.A., O'Shields, T.C. & Perlmutter, D.G. 1991. Observations on the behaviour
 of the kleptoparasitic spider, *Mysmenopsis furtive* (Araneae, Myseminidea). *Journal of* Arachnology 19: 62-66.
- Eaton, M.A., Aebischer, N.J., Brown, A.F., Hearn, R., Lock, L., Noble, D.G.,
 Musgrave, A.J., Noble, D.G., Stroud, D. & Gregory, R.D. 2015. Birds of conservation
 concern 4. *British Birds* 108: 708-746.
- 375 Ens, B.J. & Cayford, J.T. 1996. Feeding with other oystercatchers. In J.D. Goss376 Custard (ed). *The Oystercatcher: From Individuals to Populations* pp. 77- 104. Oxford
 377 University Press, Oxford.
- 378 Erlandsson, A. 1988. Food sharing vs monopolising prey: a form of kleptoparasitism in
 379 Velia Caprai (Heteroptera). *Oikos 53*: 203-206.
- 380 Giraldeau, L.-A., & Dubois, F. 2008. Social foraging and the study of exploitative
 381 behaviour. Advances in the Study of Behavior 38: 59-104.
- 382 Giraldeau, L-A. & Caraco, T. 2000. Social Foraging Theory. Princeton University Press,
 383 Princeton.
- Hatch, J.J. 1970. Predation and piracy by gulls at a ternery in Maine. *The Auk* 87: 244254.

Hamilton, I.M. 2002. Kleptoparasitism and the distribution of unequal competitors.
Behavioural Ecology 13: 260-267.

Höner, O.P., Wachter, B., East, M.L. & Hofer, H. 2002. The response of spotted
hyaenas to long-term changes in prey populations: functional response and interspecific
kleptoparasitism. *Journal of Animal Ecology* 71: 236–246.

Hopkins, C.D. & Wiley, R.H. 1972. Food parasitism and competition in two terns. *The Auk* 89: 583-594.

King, A.J., Isaac, N.J.B. & Cowlishaw, G. 2009. Ecological, social and reproductive
factors shape producer-scrounger dynamics in baboons. *Behavioural Ecology* 20: 10391049.

Leeman, L.W., Colwell, M.A., Leeman, T.S. & Mathis, R.L. 2001. Diets, energy
intake, and kleptoparasitism of nonbreeding Long-billed Curlews in a northern California
estuary. *Wilson Bulletin* 113: 194-201.

Madden, B. and Newton, S.F. 2004. Herring Gull *Larus argentatus*. In: P.I. Mitchell,
S.F. Newton, N. Ratcliffe, and T.E. Dunn, (eds.). *Seabird Populations of Britain and Ireland* pp 242-262. T. & A.D. Poyser, London.

402 Malling Olsen, K. & Larsson, H. 2003. *Gulls of Europe, Asia, and North America*.
403 Chrisopher Helm, London.

Maniscalco, J.M. & Ostrand, W.D. 1997. Seabird behaviors at forage fish schools in
Prince William Sound, Alaska. In *Forage Fishes in Marine Ecosystems*. *Proceedings of the International Symposium on the Role of Forage Fishes in Marine Ecosystems*. Alaska Sea
Grant College Program Report No. 97-01 pp 175-190. University of Alaska Fairbanks.

408 Marzluff, J.M., Bowman, R. & Donnelly, R. 2001. Avian Ecology and Conservation in
 409 an Urbanising World. Kluwer Academic Publishers, Norwell, USA.

410 Pons, J.-M., Hassanin, A. & Crochet, P.-A. 2005. Phylogenetic relationships within
411 the Laridae (Charadriiformes:Aves) inferred from mitochondrial markers. Molecular
412 Phylogenetics and Evolution 37: 686-699.

413 Proctor, N.S. & Lynch, P.J. 1993. Manual of Ornithology: Avian Structure and
414 Function. Yale, New Haven, USA.

415 **R Core Team.** 2016. R: A language and environment for statistical computing. R
416 Foundation for Statistical Computing, Vienna, Austria. URL: <u>https://www.R-project.org/</u>

- 417 **Reid, J.B.** 2004. Great Black-backed Gull *Larus marinus*. Pp 263-276 in Mitchell, P.I.,
 418 Newton, S.F., Ratcliffe, N. and Dunn, T.E. (eds.). *Seabird Populations of Britain and*419 *Ireland*. T. & A.D. Poyser, London.
- 420 Rock, P. 2005. Urban gulls: problems and solutions, *British Birds* 98: 338-355.
- 421 Schoe, M., De Iongh, H.H. & Croes, B.M. 2009. Humans displacing lions and stealing
 422 their food in Bénoué National Park, North Cameroon. *African Journal of Ecology* 47: 445423 447.
- Sol, D., Duncan, R.P., Blackburn, T.M., Cassey, P. & Lefebvre, L. 2005. Big brains,
 enhanced cognition, and response of birds to novel environments. *Proceedings of the National Academy of Science*. 102: 5460–5465.
- Sol, D., Bacher, S., Reader, S.M. & Lefebvre, L. 2008. Brain size predicts the success
 of mammal species introduced into novel environments. *American Naturalist* 172: 63–
 71.
- **Thompson, D.B.A.** 1986. The economics of kleptoparasitism: optimal foraging, host
 and prey selection by gulls. *Animal Behaviour* **34**: 1189-1205.
- 432 Wickham, H. 2009. ggplot2: Elegant Graphics for Data Analysis. Springer-Verlag, New433 York.

434 Tables

Table 1. Data used to obtain a standardised bill length measurement. Calculated by taking the mean bill length for both sexes for each species (from Malling Olsen & Larsson 2003), and then taking the mean of all four species' means. This averaging across species returned a figure of 45mm. This was closest to the bill length of Herring Gull (difference of 7mm). The Herring Gull bill length was therefore used to assess the size of food items in the field.

S	m bill	bill	species bill	S
	length (mm)	length (mm)	length (mm)	bill length
				(mm)
Common Gull	36.10	32.70	34.40	11
Black-headed Gull	33.60	31.60	32.60	12
Herring Gull	55.20	49.70	52.45	-7
Great Black-backed Gull	63.10	57.60	60.35	-15

- 445 446 **Table 2.** Summary of generalised linear models exploring the effects on log_n kleptoparasitism rate (events per
- minute) of site (with Billingsgate as reference), standardised population density, standardised prey size,
- species of kleptoparasite and season.

Model	terms	coefficients	AICc	∆ AICc
a. Main effects model	all		1474.182	6.687
b. Minimal main effects model	site	-0.86128	1468.817	1.322
	population density	+0.45983		
	prey size	+0.08091		
c. Interactions model (two-way	site	-1.35673	1467.495	0
interactions)	population density	+0.43881		
	prey size	+0.08240		

Table 3. Differences in prey size and population density between Billingsgate (N = 593) and Brancaster (N = 86). Differences between site were tested using Mann-Whitney tests.

Site	median	range	statistic		
a) Population density (birds.km ⁻²)					
Billingsgate	1.730	0.380 - 4.330	U = 50998, P < 0.0001		
Brancaster	0.0190	0.0025 - 0.1250			
b) Prey size (bill lengths)					
Billingsgate	8.00	0.75 - 20.00	U = 46575, P < 0.0001		
Brancaster	0.75	0.50 - 6.00	-		

457 Figure Legends

Figure 1. Population density plotted against standardized log_n kleptoparasitism rate at both sites, with fitted regression line and shading indicating 95% confidence intervals.

Figure 2. Prey size plotted against standardized log_n kleptoparasitism rate at both sites, with fitted regression 461 line and shading indicating 95% confidence intervals.